

# A social–ecological approach to conservation planning: embedding social considerations

Natalie C Ban<sup>1\*</sup>, Morena Mills<sup>1,2</sup>, Jordan Tam<sup>3</sup>, Christina C Hicks<sup>1</sup>, Sarah Klain<sup>3</sup>, Natalie Stoeckl<sup>4</sup>, Madeleine C Bottrill<sup>5,6</sup>, Jordan Levine<sup>3</sup>, Robert L Pressey<sup>1</sup>, Terre Satterfield<sup>3</sup>, and Kai MA Chan<sup>3</sup>

Many conservation plans remain unimplemented, in part because of insufficient consideration of the social processes that influence conservation decisions. Including social considerations in conservation planning but without integrating them with an understanding of the ecology of the region can result in an incomplete conservation approach. We suggest that linking conservation planning to a social–ecological systems (SES) framework can lead to a more thorough understanding of human–environment interactions and more effective integration of social considerations. By characterizing SES as a set of subsystems, and their interactions with each other and with external factors, the SES framework can improve our understanding of the linkages between social and ecological influences on the environment. Using this framework can help to identify socially and ecologically focused conservation actions that will benefit ecosystems and human communities, and assist in the development of more consistent evidence for evaluating conservation actions by comparing conservation case studies.

*Front Ecol Environ* 2012; doi:10.1890/110205

Despite use of the best available biophysical information and the investment of considerable time and effort, many conservation initiatives have been ineffective in motivating and guiding communities to implement the desired actions. As a result, these efforts fail to achieve their objectives (Wilson *et al.* 2007; Knight *et al.* 2008). Conservation planning is the development of spatial plans and the implementation and continued application of conservation actions (eg protection of sensitive or biodiverse areas, management of invasive species, restoration of degraded landscapes) in specific areas, usually at a regional

scale (eg encompassing a network of sites where conservation actions are undertaken, within or among ecoregions; Margules and Pressey 2000). Such planning is intended to reduce biodiversity declines in a transparent and socially responsible manner by explicitly stating overall goals and specific objectives, and then providing options for achieving them, despite limited financial resources (Margules and Pressey 2000; Pressey and Bottrill 2009). However, when such planning fails to characterize the inevitable hard choices and trade-offs involved in applying conservation actions, failures in implementation often result. Insufficient consideration of social processes (the dynamic interactions between individuals, institutions, social organizations, and cultural norms) in the social–ecological systems (SESs) in which the planning has occurred contributes substantially to this failure. Examples of contributing factors that lead to failures to implement conservation include poor understanding of the socioeconomic constraints and opportunities that shape implementation (Cowling and Wilhelm-Rechman 2007; Knight and Cowling 2007); outside agendas that conflict with local needs (Chan *et al.* 2007; Smith *et al.* 2009); and insufficient training and incentives for researchers to turn regional conservation designs into actions on the ground (Knight *et al.* 2008; Arlettaz *et al.* 2010).

Ecologists have been reluctant to engage in the messy and complex social and political aspects of implementation (Sayer *et al.* 2008), perhaps because conservation planning emerged from the natural sciences and remains predominantly rooted therein (Knight *et al.* 2006). Yet such planning – which explicitly values biodiversity – is part of a social process. It is also negatively affected by differences in power between those who make decisions about biodiversity management and those who are

## In a nutshell:

- Effective conservation planning must include both social and ecological considerations
- The social–ecological systems (SES) framework described here provides a basis for comparing conservation case studies
- A SES approach helps to explicitly consider trade-offs between ecological and social components of a system, allowing compromises to be identified
- A suite of methods and tools borrowed from the social sciences can help conservation planners understand and navigate the social complexities that underlie conservation decisions

<sup>1</sup>Australian Research Council Centre of Excellence for Coral Reef Studies, James Cook University, Townsville, Australia \*(natalie.ban@jcu.edu.au); <sup>2</sup>Global Change Institute, University of Queensland, St Lucia, Australia; <sup>3</sup>Institute for Resources, Environment and Sustainability, University of British Columbia, Vancouver, Canada; <sup>4</sup>The Cairns Institute and School of Business, James Cook University, Townsville, Australia; <sup>5</sup>Australian Research Council Centre of Excellence for Environmental Decisions, University of Queensland, St Lucia, Australia; <sup>6</sup>Conservation International, Arlington, VA

affected by its outcomes (Knight *et al.* 2008). Furthermore, conservation planning is extremely complicated because it presents many problems for which the “solutions” generally lead to the emergence of additional issues (Rittel and Webber 1973). Accordingly, while conservation plans cannot be expected to routinely produce win–win solutions, they could yield outcomes that are preferable to the various sectors involved (White *et al.* 2012), and which are better able to guide day-to-day conservation decisions. Thinking about trade-offs in terms of both the social and ecological implications of conservation actions can allow thoughtful and constructive compromises to emerge.

In this review, we discuss various ways in which social considerations have been included in conservation planning to date and offer some improvements. We suggest that linking conservation planning with an interdisciplinary SES framework (Ostrom 2009, 2010; hereafter referred to as “Ostrom’s SES framework”) is a way to effectively embed social considerations therein by broadening the predominantly ecological context to a social–ecological one.

### ■ Rationale for including social considerations in conservation planning

The rationale for integrating social considerations into conservation planning – from the perspective of planners – is that the actions that emerge are more likely to achieve their goals and to be more sustainable. Tailoring plans to the attitudes, preferences, and behaviors of stakeholders, how these change, and the process of engaging stakeholders transparently (eg engaging them throughout the planning process and providing them with all available information so that they can make informed decisions) should improve the plans overall and increase compliance with any associated recommendations (Ban *et al.* 2009). Explicitly including social considerations also creates the opportunity for planning processes to become more realistic and inclusive, clarifying the hard choices and complex trade-offs between and within conservation and other objectives (eg livelihoods and equity; Hirsch *et al.* 2010; McShane *et al.* 2011). Although an ethical imperative to include social considerations may not always be evident, because conservation is motivated by biodiversity values, conservation planners have an ethical responsibility to respect the right of local communities to be an integral part of the planning process.

### ■ Social considerations in conservation planning to date

A number of approaches have been suggested for expanding the inclusion of social considerations in conservation planning. Knight *et al.* (2006) outlined an operational model for this that emphasized the need for implementation strategies and the importance of including stakeholders throughout the process, to reflect local knowledge when gathering information about the region under consideration.

Similarly, the stages involved in systematic conservation planning have been increased from the initial ecology-centric approach articulated by Margules and Pressey (2000) to include five new stages (Pressey and Bottrill 2009), most of which concern the social, economic, and political context in which the conservation initiative will take place (Table 1; WebFigure 1). While current conservation planning frameworks (Knight *et al.* 2006; Pressey and Bottrill 2009) include some social considerations, they do not yet provide a truly integrative approach that recognizes substantial social processes and social–ecological linkages.

Practical developments that have occurred as a result of calls to improve the integration of social considerations into conservation planning fall into two categories, both of which are linked to the assessment part of conservation planning: (1) use of spatial data pertaining to existing resource use, and (2) the addition of social assessments, including identification of areas where conservation is more likely to succeed (ie “conservation opportunities”).

### *Spatial data on human uses*

Spatial data on how people use resources are increasingly being incorporated into conservation assessments; these are usually represented as threats to biodiversity or as costs associated with conservation actions. For instance, when human activities represent a threat to biodiversity (eg land clearing), planners either avoid highly threatened areas (eg areas slated for land clearing) to minimize conflict (when other, less sensitive areas exist that have the same biodiversity values) or give priority to areas of high biodiversity value that are highly threatened to protect them before land clearing occurs (where there are no viable alternatives) (Pressey and Taffs 2001). Planners deal with costs in similar ways, where the “costs” relate to acquisition (eg land value; Ando *et al.* 1998; Carwardine *et al.* 2010), management, damage to an ecosystem, and loss of extractive opportunities (for a review of terrestrial and marine systems respectively, see Naidoo *et al.* 2006; Ban and Klein 2009). The term “costs” is also used more generically, to refer to liabilities related to past or present human uses. Although there is scope for improving the representation of costs in conservation planning, once the costs of a planning process are defined, planners commonly use decision support tools (such as Marxan) to minimize costs while achieving conservation objectives. For example, the zoning of the Great Barrier Reef in Australia sought to represent at least 20% of the area of each bioregion while reducing human impacts (Fernandes *et al.* 2005).

### *Social assessments*

Social assessments (also termed situation analyses, social analyses, or stakeholder assessments) are a common component of conservation planning. These assessments contextualize aspects of the social systems that exist in the planning region, describing the social, cultural, economic,

and political conditions in the area (Knight *et al.* 2006; Conservation Measures Partnership 2007; Cowling and Wilhelm-Rechman 2007; Figure 2, stages 1–3, 5). Some social assessments focus on the local opportunities for conservation that emerge where social factors align to create a willingness among community stakeholders to implement conservation actions; these may be community-led initiatives or may be linked to regional planning (Cowling and Wilhelm-Rechman 2007). Game *et al.* (2011) provide an example of the latter approach, working with communities in the Solomon Islands to identify protected areas through several rounds of meetings between community members and conservation planners, thereby combining the priorities of the communities with a systematic assessment of areas characterized by a high level of biodiversity.

### **Limitations of current approaches**

To date, the inclusion of social considerations in conservation planning has been limited in several ways. First, although social data (eg opportunity costs) have increasingly been included in conservation assessments, the approaches used have not been consistent (Timko and Satterfield 2008). Gaining an understanding of what kinds of social factors matter, why they matter, and how this information should be collected, integrated, and interpreted has proved challenging. Second, incorporating social data into conservation assessments requires these data to be simplified and mapped, when in fact some social or cultural priorities may be distinctly aspatial and/or that information was not articulated in spatial terms initially. For example, many ethnographic insights into the underlying tensions that influence compliance (Fabinyi 2010) were not intended to be mapped and so lack spatial representation, making it difficult to include them in conservation assessments. Third, the process and products of systematic conservation planning tend to be static, prescriptive, and often technical, which limits the scope for including dynamics, values, and trade-offs among different objectives when these considerations are not articulated in prescriptive and technical (usually “measurable”) terms. Fourth, limited guidance exists on how to move from recognition of the need to address social aspects of resource use in conservation planning to actually incorporating these into planning. Despite these limitations, the advantage of the systematic conservation planning framework (Margules and Pressey 2000) is that it is transparent and has been widely used. It can serve as a starting point to infuse a more comprehensive view of social considerations and trade-offs into conservation (see also Satterfield *et al.* in press).

### **■ Insights from the social sciences**

#### **Critical perspectives**

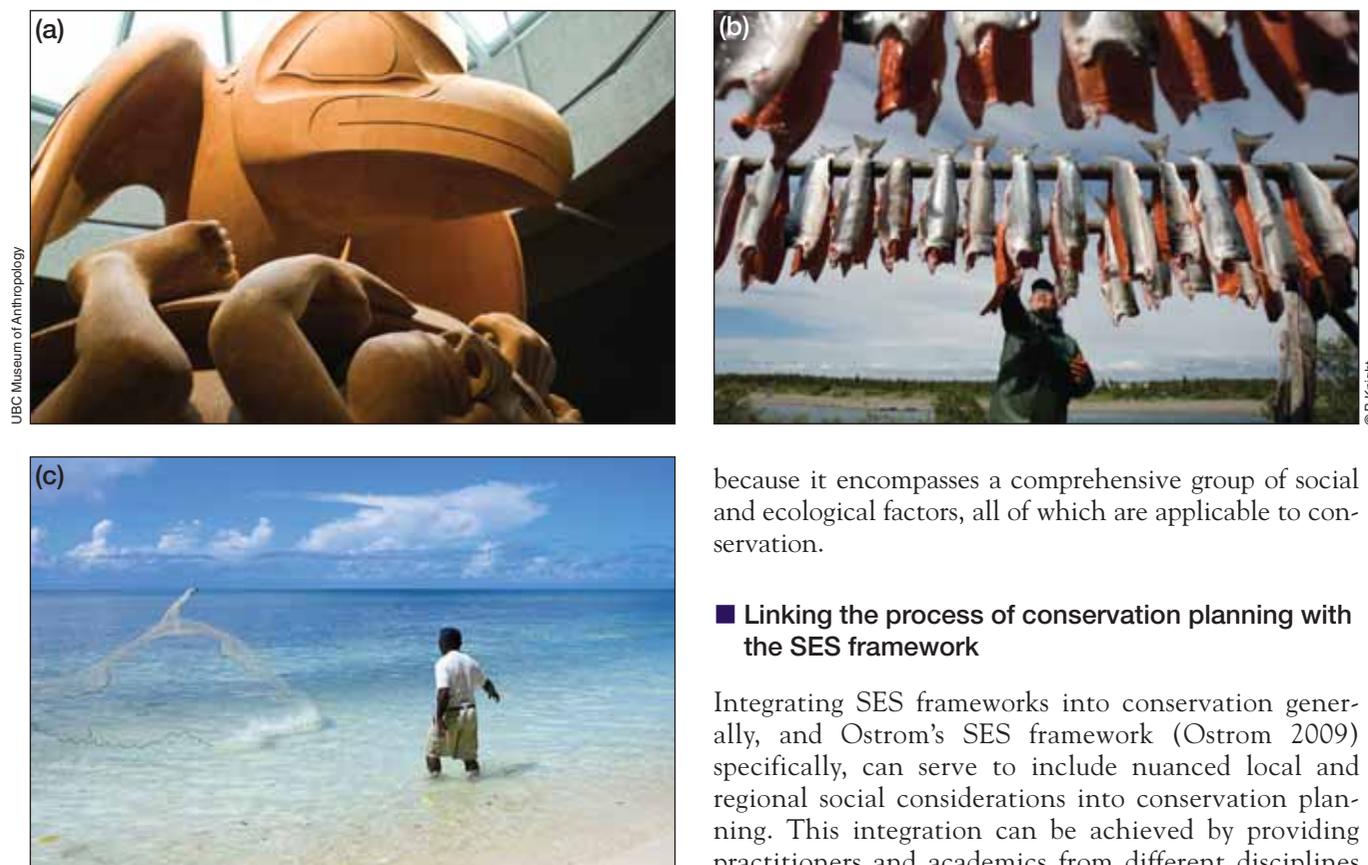
Given that critical analysis is central to many of the social sciences, these disciplines offer fertile territory for a

detailed scrutiny of conservation practices. For example, there have been in-depth criticisms of various aspects of conservation (eg exclusion or eviction of local peoples) and community-level conflicts (Agrawal and Gibson 1999; Brosius 1999; Brechin *et al.* 2003). Such criticisms should persuade conservation planners to take into account alternative viewpoints (eg whether biodiversity matters in its own right, or whether meeting basic human needs should be a priority) and to bear in mind how a wide spectrum of different values can influence choices; such considerations help to clarify some of the motives underlying human behavior. However, there is understandable wariness among conservation planners of such critical perspectives. The exhaustive analysis of social systems, such as in anthropological studies, could in principle greatly enhance the appropriateness, and effectiveness, and subsequent implementation of conservation plans (Harper 2002). The challenge remains: to increase engagement and communication between natural and social scientists so as to improve conservation in practice (Igoe 2011; Redford 2011), in part by providing a common language and framework for various fields to contribute to a fuller understanding of the drivers and impacts of conservation initiatives (Ostrom 2009).

#### **Interactions between people and the environment**

Insights and techniques borrowed from the social sciences have the potential to create more realistic expectations regarding the outcomes of conservation initiatives and to enhance the effectiveness, efficiency, and sustainability of these initiatives, by providing a better understanding of the complex linkages between people and the environment at multiple scales (Figure 1). These linkages have been investigated by several communities of scholars, including political and human ecologists, ecological anthropologists, and economists. The study of SESs draws insights from all of these fields, and others as well (Berkes *et al.* 2003). A deeper understanding of SES dynamics can highlight multiple issues that are relevant to conservation. For example, SES studies can help explain the benefits and drawbacks of multiple knowledge systems (eg different ways of viewing the world), informal institutions (eg the rules that people abide by, including social norms), and cross-scale networks. Most importantly, a SES view emphasizes the unpredictable, dynamic, and evolved nature of linked social and ecological systems (Berkes *et al.* 2003).

Ostrom’s SES framework was developed to provide an understanding of the processes that lead to improvements in or deterioration of renewable natural resources (Ostrom 2009). The framework grew out of a large body of interdisciplinary research about coordinated resource management successes and failures. It divides SESs into subsystems, based on the resource (eg forests, coastal areas, etc), resource units (eg trees, fish), governance systems (eg management of a forest or a coastal area), and



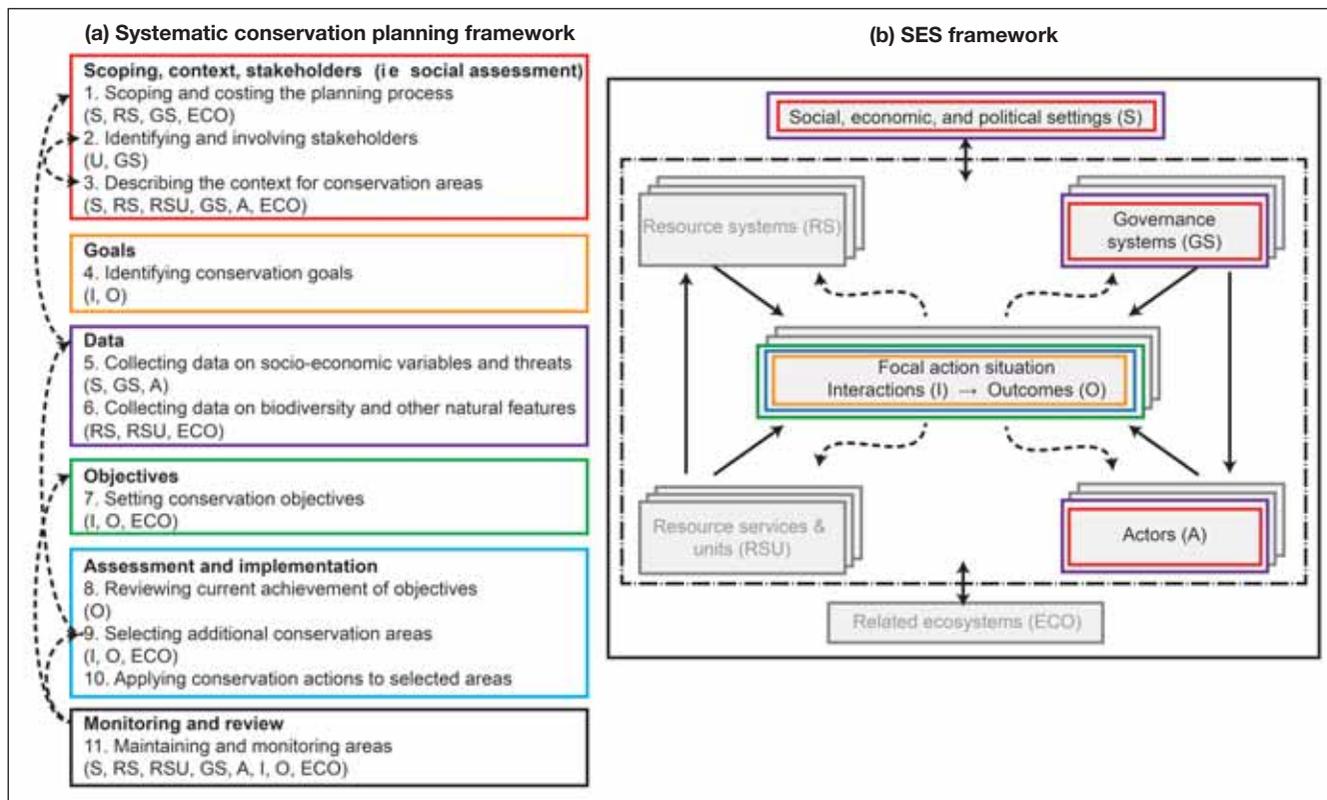
**Figure 1.** Cultural linkages between people and the environment highlight the importance of considering both ecological and social aspects in conservation planning. (a) The creation stories of many different cultures link people and animals, such as the Haida First Nation’s legend of the raven and first humans as depicted in this sculpture by Bill Reid, making it inappropriate to ignore that linkage. Understanding food and harvesting traditions, such as (b) smoking salmon in Alaska and (c) catching nearshore reef fish in Palau, helps conservation planning meet subsistence needs.

actors (ie stakeholders, such as hikers, loggers, and fishers). These four subsystems interact with each other and with the overarching social, economic, and political settings and related ecosystems (ie interactions and outcomes, with variables like harvesting levels, deliberative processes, activities carried out by communities, and social and ecological performance measures), leading to resource management outcomes (Ostrom 2009). While the framework emerged mainly from studies at local scales, it is just as applicable at regional and even global scales. Other frameworks (principally ones that are not connected with conservation) that have similarly linked social and ecological components include the following: “pressure–state–impacts–response” (Turner 2000), “sustainable rural livelihoods” (Scoones 1998), “disaster resilience of place” (Cutter *et al.* 2008), and “pressure and release” (Blaikie *et al.* 1994). However, we contend that the SES framework, as presented by Ostrom (2009), is the most appropriate for use in conservation planning

because it encompasses a comprehensive group of social and ecological factors, all of which are applicable to conservation.

#### ■ Linking the process of conservation planning with the SES framework

Integrating SES frameworks into conservation generally, and Ostrom’s SES framework (Ostrom 2009) specifically, can serve to include nuanced local and regional social considerations into conservation planning. This integration can be achieved by providing practitioners and academics from different disciplines with a common vocabulary and a logical structure for classifying factors deemed as important influences in developing and implementing a conservation plan. In particular, linking the stages of systematic conservation planning (Pressey and Bottrill 2009) to Ostrom’s SES framework (Ostrom 2009) (Figure 2; see WebFigures 1 and 2 for more detail) allows planners to think beyond the usual concerns. For example, at present, conservation planners do not routinely consider existing or potential governance systems, even though this might aid conservation. By understanding governance variables (ie the attributes that constitute a governance system, such as rule-making organizations, social norms that determine informal procedures for management, collaborations between different agencies or sectors, legal systems, current rules, and policy tools), planners can expand their approach and/or conservation actions to consider some or all of these factors. Thinking about such variables will also highlight the importance of multiple scales (eg the dynamics of government agencies and non-governmental groups at levels above and below the scale at which planning is occurring). Furthermore, one area where alternative theories, perspectives, and values can be integrated into conservation plans is within the “action situation” in Ostrom’s SES framework (Ostrom 2010). The action situation is a step in the planning process whereby proposed conservation actions and their likely outcomes can be evaluated in light of the stakeholders’ opinions and beliefs. In this way, Ostrom’s SES framework can provide a first



**Figure 2.** Linkages between frameworks for (a) systematic conservation planning (Pressey and Bottrill 2009) and (b) social–ecological systems (SES) (Ostrom 2007, 2009, 2010). In (a), the colors depict categories of stages: dashed arrows represent feedback loops between stages, or, where arrows enter boxes, feedback loops between specific stages. The colors in (b) link the most directly relevant social–ecological subsystems to planning stages in (a). In (b), grey text relates primarily to ecological considerations, which we do not discuss; solid arrows indicate direct links, and dashed arrows indicate feedbacks; the dashed box denotes the action situation embedded in a broader SES; the multiple versions of boxes symbolize that there can be multiple subsystems for each action situation. Abbreviations after stages in (a) refer to those components of the SES, defined in (b), which are relevant to specific stages of systematic conservation planning. See WebFigures 1 and 2 for more detail.

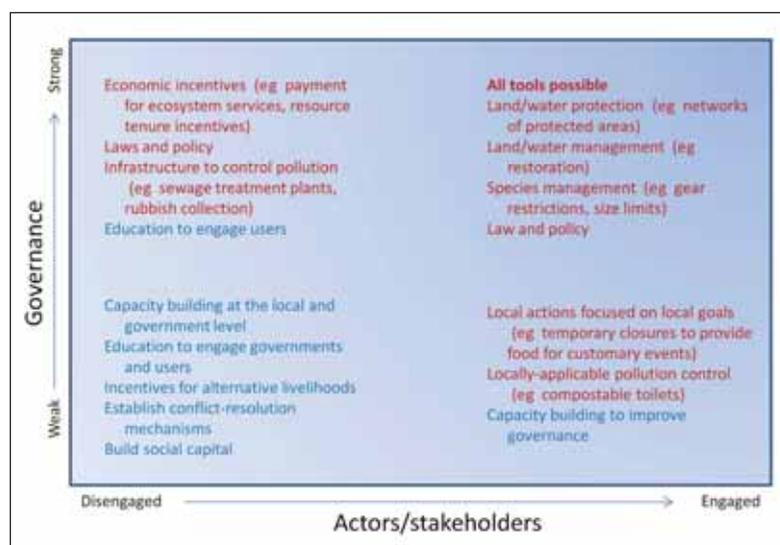
step toward bringing together social considerations and conservation practices.

In practice, designing a conservation plan that incorporates SES thinking might entail some or all of the following: (1) careful consideration of both the social and ecological elements of the target region and their interactions; (2) inclusion of social scientists familiar with SES ideas in the planning process; (3) building capacity within existing complementary local or regional institutions; (4) discussion among planners about which methods, tools, and data are appropriate to the planning context (WebTable 1); and (5) prioritizing data collection and analysis. Furthermore, mainstreaming key social concepts and methods into existing planning frameworks through policies and day-to-day activities can promote acceptance and long-term commitment of social considerations. Admittedly, linking the frameworks will take time, experimentation, and a commitment by planners. We therefore highlight two specific ways in which conservation planning and SES thinking can be linked: (1) Ostrom's SES framework can be used to guide comparative analyses of conservation case studies, thereby creating an evi-

dence-base for conservation actions; and (2) social objectives and stakeholders' goals should be incorporated into conservation planning.

### Comparative analysis of conservation case studies

The conservation community currently lacks a general understanding of which conservation actions work in different contexts (Sutherland *et al.* 2004). Assessment of conservation cases under a common framework could help to develop an understanding of the relationships between social contexts and effective conservation actions. This in turn would help identify context-appropriate conservation actions, thereby improving the chances of successful conservation outcomes. A relatively modest allocation of funding to monitor progress could help to assess the social and ecological effectiveness of particular strategies (Neugarten *et al.* 2011). Although conservation planners could carry out such an assessment as part of the planning process (ie assess which conservation actions are more likely to be effective, based on past experiences), academics could also play a role here.



**Figure 3.** Theoretical model relating conservation and supporting actions to variation in engagement of stakeholders and strength of governance. Planners face two options: select direct conservation actions based on the existing context (red text), and/or select supporting actions to change the context (blue text; ie to better engage users and/or to strengthen governance). The actions shown here are illustrative; appropriate specific actions will vary according to the context. The actions are not intended to be mutually exclusive and multiple actions might be necessary in one place. Schematic based on a similar concept by McClanahan *et al.* (2008); actions are loosely based on Salafsky *et al.* (2008).

In a SES, conservation actions are immediately embedded in a complex web of social and ecological processes and interactions. Ostrom's SES framework could cue planners to potentially relevant variables, which would help determine how the various actors should be involved and what constitutes an appropriate conservation action. For instance, a particular form of protected area might require the involvement of stakeholders who have strong leadership roles, traditional knowledge, and social capital, as well as a need for a relatively well-developed governance system (Figure 3). Where governance is limited, conservation actions might need to be altered, or may need to be coupled with or preceded by supporting actions that bolster local institutions and governance. These arrangements could then be incorporated into the planning process (eg in this example, stages 3 and 10 of Figure 2a).

In work that led up to the SES framework, Ostrom (1990) identified eight principles (more recently reformulated to 11; Cox *et al.* 2010) that, in different combinations, are commonly found in institutions that successfully manage resources. These principles (see WebTable 2) describe under what conditions trust and reciprocity are anchored to local norms. These principles can be used as a starting point to investigate systems of conservation interest and identify appropriate and feasible conservation actions. Planners should seek conservation opportunities in situations where these principles are already in operation. Furthermore, Ostrom's SES framework can be used as a diagnostic tool to identify

additional design principles that relate to both social and ecological variables.

### ■ Social goals, objectives, and ecosystem services

Conservation planners are actively pursuing practical ways to integrate social considerations into conservation initiatives, including through social goals, objectives, and ecosystem services, potentially implemented through zoning. Social goals are general statements regarding desirable outcomes that might support conservation (eg “reduce local poverty”; Holland *et al.* 2009), but are not direct conservation goals (a direct conservation goal, in contrast, might be to protect biodiversity in perpetuity). Social objectives are specific and quantitative articulations of goals (eg “expand available schooling to 85% of the population under the age of 16”). Some social goals and objectives could serve to increase compliance with the rules governing conservation areas (eg a goal might be: “to ensure viable fisheries livelihoods”; the objective would be: “fishing-designated areas provide at least 90% of previous catch for each fishery”).

SES thinking can help planners identify social objectives that may not be immediately obvious, such as building trust and forging links within the social network structure. Multiple objectives can be pursued through zoning – that is, geographically defining areas where different types of activities are allowed (eg marine spatial planning in the oceans; Ehler 2008; Klein *et al.* 2010). Similarly, planners are starting to prioritize management actions to provide ecosystem services that benefit people and biodiversity (Chan *et al.* 2006; Cowling *et al.* 2008), thereby linking conservation goals and social concerns. Incorporating multiple objectives is extremely difficult, however, and although decision support tools can be helpful, they are merely tools and cannot provide satisfying outcomes in and of themselves. Rather, they may provide options or potential scenarios that can be discussed during the planning phase.

### ■ Benefits of linking conservation planning and SES thinking

Linking SES thinking and conservation planning could highlight trade-offs among different objectives inherent in conservation decisions. Planners will always be faced with hard choices, but clearer trade-offs may compel them to modify biodiversity priorities or justify the inequalities they create. We therefore expect the emergence of a deeper understanding about how trade-offs are perceived and experienced from multiple perspectives (McShane *et al.* 2011). In the long run, this honesty and clarity may



**Figure 4.** Involving people in planning. (a) Provincial leader annotating a map to identify areas for potential future marine conservation in Fiji. (b) Village leaders deciding on their vision for future resource management practices in Ovalau, Fiji. (c) Social scientist carrying out interviews with resource users in Kenya. (d) Map with notes from an interview about marine values in British Columbia, Canada.

yield more effective conservation initiatives. Furthermore, an additional benefit of integrating SES thinking into conservation planning is based on the former's emphasis on dynamics, interactions, and processes at multiple scales, whereas planning is all too often viewed as occurring in a static context.

#### ■ Methods and tools

Linking SES thinking and conservation planning requires analytical techniques that can provide insight into the social components and interactions of SES. Connecting the two frameworks also requires integration of different methods and tools into a coherent process. The social sciences offer many methods and tools for incorporating social considerations into conservation planning (eg Figure 4), but most are rarely applied. We outline examples of some tools here, and touch on some key examples in WebPanel 1 and WebTable 1. We broadly categorize methods and tools by their value to

three interlinked facets of the planning process: (1) describing and analyzing the SES in which planning is taking place; (2) eliciting input into the planning process; and (3) analyzing and selecting appropriate conservation actions.

First, many methods from the social sciences can assist with describing and analyzing the SES in which planning is taking place. A starting point might be to investigate aspects of the social–ecological system that have been repeatedly found to be linked to successful resource management (ie design principles; WebTable 2). One example is the existence of clearly defined spatial and social boundaries. Another is the congruence between rules about extracting resources (appropriation rules), rules about labor, materials, and money for management (provision rules), and the local conditions (Ostrom 1990; Cox *et al.* 2010). Questionnaires can be designed to establish levels of agreement and clarity regarding spatial and social boundaries. “The institutional analysis and development” framework (Ostrom 1990) helps in understand-

ing existing institutional arrangements within a region; this in turn helps to avoid problems with coordination and conflict between existing and future rules governing resource use (Imperial 1999).

Second, methods from the social sciences can help to elicit input into the planning process, and finally, their methods and tools can help to select and analyze appropriate conservation actions. The design principles can again serve as a starting point; Ostrom (1990) identified critical aspects in resource governance (participation of stakeholders in rule-making, monitoring and sanctioning, and recognition of local autonomy). One example of this approach is through participation in the planning process, which can be garnered through collaborative mapping (Ban *et al.* 2008). This design principle could also be used to select appropriate actions. Questionnaires can be designed to establish current and historic levels of participation in resource management, which can assist in predicting the likelihood of stakeholder compliance. Participatory approaches (eg through structured decision making) can be used to elicit values of stakeholders toward conservation, and thus can help to identify appropriate conservation options (Gregory *et al.* 2001; see also WebTable 1).

### ■ Limitations of local and regional conservation planning

Our review has focused on regional conservation planning, but macro-level forces – such as international policies, the national economy, materialism, and water and food insecurity – strongly influence conservation endeavors at all scales. While these forces might be acknowledged, addressing them has been beyond the remit of conservation planning, and so they have remained a core challenge for conservation (not just conservation planning). Ostrom's SES framework can be a starting point for exploring these forces because in theory it can be used to investigate SES at all scales. Understanding these types of pressures also allows planners to avoid inadvertently reinforcing them or simply displacing their impacts to other areas of conservation importance. The future task of conservation planners might not only be to build resilient, adaptive, and culturally sensitive institutions for regional biodiversity protection, but also to broaden views and strategies so as to contribute to changing macro-level societal structures.

### ■ Conclusion

A major shift in the planning community will be needed to integrate social–ecological thinking into conservation planning – one that places the same importance on social considerations as on ecological ones, and that seeks to integrate the two. We hope that this shift will begin with the integration of concepts that have emerged from the SES framework into conservation planning.

Ultimately, lessons from SES thinking and research and

other insights from the social sciences can enhance conservation planning efforts, and potentially help secure a more certain future for conservation – and hence biodiversity – in a world of growing human needs and impacts.

### ■ Acknowledgements

We thank the International Council for Canadian Studies and the Government of Canada's Department of Foreign Affairs and International Trade for funding support through their International Linkages Grant, which enabled us to hold a workshop that formed the basis of this paper. NCB and RLP thank the Australian Research Council for support (grant DP1096453 for NCB). E Ostrom generously shared the latest version of the SES framework with us in 2011. She will be missed.

### ■ References

- Agrawal A and Gibson C. 1999. Enchantment and disenchantment: the role of community in natural resource conservation. *World Dev* 27: 629–49.
- Ando A, Camm J, Polasky S, and Solow A. 1998. Species distributions, land values, and efficient conservation. *Science* 279: 2126–28.
- Arlettaz R, Schaub M, Fournier J, *et al.* 2010. From publications to public actions: when conservation biologists bridge the gap between research and implementation. *BioScience* 60: 835–42.
- Ban NC and Klein CJ. 2009. Spatial socio–economic data as a cost in systematic marine conservation planning. *Conserv Lett* 2: 206–15.
- Ban NC, Picard C, and Vincent ACJ. 2008. Moving towards spatial solutions in marine conservation, with indigenous communities. *Ecol Soc* 13: 32.
- Ban NC, Picard CR, and Vincent ACJ. 2009. Comparing and integrating community-based and science-based approaches in prioritizing marine areas for protection. *Conserv Biol* 23: 899–910.
- Berkes F, Colding J, and Folke C (Eds). 2003. Navigating social–ecological systems: building resilience for complexity and change. Cambridge, UK: Cambridge University Press.
- Blaikie PM, Cannon T, Davis I, and Wisner B. 1994. At risk: natural hazards, people's vulnerability, and disasters. New York, NY: Routledge.
- Brechin SR, Wilshusen PR, Fortwangler CL, and West PC (Eds). 2003. Contested nature: promoting international biodiversity with social justice in the twenty-first century. Albany, NY: State University Press of New York.
- Brosius JP. 1999. Analyses and interventions: anthropological engagements with environmentalism. *Curr Anthropol* 40: 277–309.
- Carwardine J, Wilson KA, Hajkovicz SA, *et al.* 2010. Conservation planning when costs are uncertain. *Conserv Biol* 24: 1529–37.
- Chan KMA, Pringle RM, Ranganathan J, *et al.* 2007. When agendas collide: human welfare and biological conservation. *Conserv Biol* 21: 59–68.
- Chan KMA, Shaw MR, Cameron DR, *et al.* 2006. Conservation planning for ecosystem services. *PLoS Biol* 4: e379.
- Conservation Measures Partnership. 2007. Open standards for the practice of conservation, version 2.0. Washington, DC: Conservation Measures Partnership.
- Cowling RM, Egoh B, Knight AT, *et al.* 2008. An operational model for mainstreaming ecosystem services for implementation. *P Natl Acad Sci USA* 105: 9483.
- Cowling RM and Wilhelm-Rechman A. 2007. Social assessment as

- a key to conservation success. *Oryx* **41**: 135–36.
- Cox M, Arnold G, and Tomás SV. 2010. A review of design principles for community-based natural resource management. *Ecol Soc* **15**: 38.
- Cutter SL, Barnes L, Berry M, *et al.* 2008. A place-based model for understanding community resilience to natural disasters. *Glob Environ Chang* **18**: 598–606.
- Ehler C. 2008. Conclusions: benefits, lessons learned, and future challenges of marine spatial planning. *Mar Policy* **32**: 840–43.
- Fabinyi M, Knudsen M, and Segi S. 2010. Social complexity, ethnography and coastal resource management in the Philippines. *Coast Manage* **38**: 617–32.
- Fernandes L, Day J, Lewis A, *et al.* 2005. Establishing representative no-take areas in the Great Barrier Reef: large-scale implementation of theory on marine protected areas. *Conserv Biol* **19**: 1733–44.
- Game ET, Lipsett-Moore G, Hamilton R, *et al.* 2011. Informed opportunism for conservation planning in the Solomon Islands. *Conserv Lett* **4**: 38–46.
- Gregory R, McDaniels T, and Fields D. 2001. Decision aiding, not dispute resolution: creating insights through structured environmental decisions. *J Policy Anal Manag* **20**: 415–32.
- Harper J. 2002. Endangered species: health, illness and death among Madagascar's People of the Forest. Durham, NC: Carolina Academic Press.
- Hirsch PD, Adams WM, Brosius JP, *et al.* 2010. Acknowledging conservation trade offs and embracing complexity. *Conserv Biol* **25**: 259–64.
- Holland TG, Peterson GD, and Gonzalez A. 2009. A cross-national analysis of how economic inequality predicts biodiversity loss. *Conserv Biol* **23**: 1304–13.
- Igoe J. 2011. Rereading conservation critique: a response to Redford. *Oryx* **45**: 333–34.
- Imperial MT. 1999. Institutional analysis and ecosystem-based management: the institutional analysis and development framework. *Environ Manage* **24**: 449–65.
- Klein CJ, Steinback C, Watts M, *et al.* 2010. Spatial marine zoning for fisheries and conservation. *Front Ecol Environ* **8**: 349–53.
- Knight AT, Cowling RM, Rouget M, *et al.* 2008. Knowing but not doing: selecting priority conservation areas and the research–implementation gap. *Conserv Biol* **22**: 610–17.
- Knight AT and Cowling RM. 2007. Embracing opportunism in the selection of priority conservation areas. *Conserv Biol* **21**: 1124–26.
- Knight AT, Cowling RM, and Campbell BM. 2006. An operational model for implementing conservation action. *Conserv Biol* **20**: 739–50.
- Margules CR and Pressey RL. 2000. Systematic conservation planning. *Nature* **405**: 243–53.
- McShane TO, Hirsch PD, Trung TC, *et al.* 2011. Hard choices: making trade-offs between biodiversity conservation and human well-being. *Biol Conserv* **144**: 966–72.
- Naidoo R, Balmford A, Ferraro PJ, *et al.* 2006. Integrating economic costs into conservation planning. *Trends Ecol Evol* **21**: 681–87.
- Neugarten RA, Wolf SA, Stedman RC, and Tear TH. 2011. Integrating ecological and socioeconomic monitoring of working forests. *BioScience* **61**: 631–37.
- Ostrom E. 1990. Governing the commons: the evolution of institutions for collective action. New York, NY: Cambridge University Press.
- Ostrom E. 2009. A general framework for analyzing sustainability of social–ecological systems. *Science* **325**: 419–22.
- Ostrom E. 2010. Beyond markets and states: polycentric governance of complex economic systems. *Am Econ Rev* **100**: 641–72.
- Pressey RL and Bottrill MC. 2009. Approaches to landscape- and seascape-scale conservation planning: convergence, contrasts and challenges. *Oryx* **43**: 464–75.
- Pressey RL and Taffs KH. 2001. Scheduling conservation action in production landscapes: priority areas in western New South Wales defined by irreplaceability and vulnerability to vegetation loss. *Biol Conserv* **100**: 355–76.
- Redford KH. 2011. Misreading the conservation landscape. *Oryx* **45**: 324–30.
- Rittel HWJ and Webber MM. 1973. Dilemmas in a general theory of planning. *Policy Sci* **4**: 155–69.
- Satterfield T, Gregory J, Klain S, *et al.* Culture, intangibles and metrics in environmental management. *J Environ Manage*. In press.
- Sayer J, Bull G, and Elliott C. 2008. Mediating forest transitions: “grand design” or “muddling through”. *Conserv Soc* **6**: 320.
- Scoones I. 1998. Sustainable rural livelihoods: a framework for analysis. Brighton, UK: Institute of Development Studies.
- Smith RJ, Verissimo D, Leader-Williams N, *et al.* 2009. Let the locals lead. *Nature* **462**: 280–81.
- Sutherland WJ, Pullin AS, Dolman PM, and Knight TM. 2004. The need for evidence-based conservation. *Trends Ecol Evol* **19**: 305–08.
- Timko J and Satterfield T. 2008. Seeking social equity in national parks: experiments with evaluation in Canada and South Africa. *Conserv Soc* **6**: 238–54.
- Turner RK. 2000. Integrating natural and socio–economic science in coastal management. *J Marine Syst* **25**: 447–60.
- White C, Halpern BS, and Kappel CV. 2012. Ecosystem service tradeoff analysis reveals the value of marine spatial planning for multiple ocean uses. *P Natl Acad Sci USA* **109**: 4696–4701.
- Wilson KA, Underwood EC, Morrison SA, *et al.* 2007. Conserving biodiversity efficiently: what to do, where, and when. *PLoS Biol* **5**: e223.